

Activation energy, Temperature Coefficient and Q₁₀ Value Estimations of the Growth of *Serratia marcescens* strain DRY6 on SDS

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ABSTRACT

Sodium dodecyl sulfate, occasionally known as SDS, is a widely used anionic surfactant globally as a foaming component in a wide variety of cleaning products, including shampoos, toothpaste, and detergents. Large quantities of SDS are released into the environment despite the fact that they are hazardous and have the ability to create foam. This can lead to problems in sewage treatment facilities. There are a few different models that may be utilized to mimic the growth rate of microbes on a variety of different medium at various temperature. Arrhenius is one of the most often used models partly since it has a small number of parameters. Temperature frequently influences the progress and metabolic activities of microbes on the substrates they are growing on. Because of their small size, microbes are very sensitive to variations in temperature. In order to describe the development of *Serratia marcescens* strain DRY6 on SDS, a discontinuous apparent activation energy with a chevron-like graph was used, and the graph's break point was set at 28.05 degrees Celsius. Following the completion of the regression study, two activation temperatures were established: 20-27 degrees Celsius and 30-42 degrees Celsius, with respective activation energies of 41.72 and 84.72 kilojoules per mole. Within the temperature range that was taken into account, it was projected that the Q₁₀ value would be 2.905, and the theta value would be 1.11. (30-42 °C). This study is especially helpful in projecting SDS breakdown and migration during bioremediation because of its comprehensive nature.

INTRODUCTION

Detergents have been shown to be harmful to marine life. According to previous findings, anionic surfactants are toxic to a wide spectrum of aquatic creatures from 0.0025 to 300 mg/L [1]. It altered the life cycle of aquatic species and caused behavioral changes. Another study found that the oyster digestive gland is susceptible to SDS exposure, creating a detrimental disruption in the nutritional and metabolic activities of the oyster, resulting in reduced oyster survivorship [2]. As more anionic surfactants are discharged into bodies of water, pollution from these substances will increase the deleterious effects on invertebrates and crustaceans. SDS, commonly known as Sodium Lauryl Sulfate, is the most often used anionic detergent in home items such as toothpastes, shampoos, bubble baths, cosmetics, shaving foams, and detergents [3]. However, in the industry, it is employed as a wool cleaning agent, a de-inking agent in the paper industry, a leather softening agent, a penetrant, a flocculating agent, and a key component of fire-fighting equipment, engine degreasers,

floor cleaners, and car wash soaps [4–6]. The existence of SDS in the environment is mostly due to its presence in home and industrial effluents, as well as its direct discharge from some applications. Because of this, SDS remediation is critical. Microorganisms are recognized for their capacity to degrade organic substances such as SDS [7–11], and their usage as bioremediation agents is economically critical for the removal of xenobiotic pollutants. One of the first reports of an SDS-degrading bacteria was *Pseudomonas* sp. strain C12B biodegrading anionic surfactant under aerobic conditions [12].

It is not very common to find research on bacteria that display the ability to degrade numerous xenobiotics. However, because polluted sites typically include a wide variety of inorganic and organic contaminants, isolating such unique bacteria is absolutely necessary. Because of their microscopic nature, SDS decomposition is sensitive to temperature changes. The physiology of an organism may be modulated by temperature, which enables the organism to better adapt to

changing environments. When researching the process of chemical breakdown by microbes, temperature is an important factor to take into consideration. It is common practice to use the temperature function Arrhenius model to compute the apparent activation energy, H^* , which is thought to be present for either growth or decay on distinct metabolic substrates. The temperature function Arrhenius model has become increasingly popular for use in the investigation of the growth and decomposition rates of bacteria [13–21].

Throughout the majority of the temperature range, the delta H (H^*) value is nearly constant. For extreme temperatures, this value can vary by three or four times depending on the temperature range [22]. In some studies, the model may be incorrect when applied to the entire temperature range of the bacterial process [23]. When modeling temperature effects, Arrhenius' model is frequently used, but it is less frequently used when the temperature ranges are large [24]. The Arrhenius plot may also show a previously identified transition, which is a fast shift in activation energy [25]. Because Arrhenius' model includes the fewest parameters, it is pretty widely accepted by researchers [24].

Because of its biological underpinnings and non-linear behavior this rival Ratkowsky model is also based on the linear growth premise but lacks constant development [26]. This means that the Arrhenius models are essential for figuring out how temperature affects bacterial growth. The Arrhenius parameter estimate is obtained by plotting an Arrhenius plot and then performing a linear regression on the data. Arrhenius plot analysis and the effect of temperature on *Pseudomonas* sp. strain DRYJ7's SDS growth were the subject of a similar study a number of years ago [16]. This study found that a bacterium can break down SDS at a number of different activation energies, which was previously unknown and will be extremely useful in predicting how SDS breakdown and transport will occur during bioremediation.

MATERIALS AND METHODS

Growth and maintenance of bacterium

Serratia marcescens strain DRY6 growth and characterization on SDS has been published previously and stored in the university's culture collection unit [27]. The SDS as sole carbon source medium composition (g L^{-1}) was as follows; KH_2PO_4 , (1.36), KNO_3 , (0.5), MgSO_4 (0.01), CaCl_2 (0.01), Na_2HPO_4 (1.39) and $(\text{NH}_4)_2\text{SO}_4$ (7.7). The final concentration of the standard trace elements was 0.01 mg L^{-1} in the medium. Sodium dodecyl sulfate, filter-sterilized, was added to the medium as a carbon source. Sodium dodecyl sulphate, a carbon source, was added to the medium at a final concentration of 1.0 g/L via filter-sterilization (Dhouib et al. 2003). The bacterium was grown on nutrient agar plates treated with SDS at the same dose for 5 days at 30 °C. The colony count method was used to measure the growth of microorganisms.

Measurement of the Activation energy of growth on SDS

Biodegradation growth data from *Serratia marcescens* strain DRY6 previously isolated as an SDS-degrading bacterium was processed [27] by converting the temperature-dependent growth rates to natural logarithms.

The Arrhenius equation [28] is as follows,

$$\mu = Ae^{-\frac{E_a}{RT}} \quad [\text{Eqn. 1}]$$

Where R is the universal gas constant ($0.008314 \text{ kJ/molK}^{-1}$), T is the absolute temperature (Kelvin = $^{\circ}\text{C} + 273.15$), E_a is the activation energy (kJ/mol) and A physically signifies the rate constant at which all the participating molecules possess sufficient energy prior reaction ($E_a = 0$). A linearized form is given via the plot of log normal growth rate against $1/T$ and the equation is as follows;

$$\ln \mu = \ln A - \frac{E_a}{R} \cdot \frac{1}{T} \quad [\text{Eqn. 2}]$$

Coefficient of Q_{10} estimation

The Q_{10} value is estimated via the following equation.

$$Q_{10} = e^{\left(\frac{E_a}{R}\right)\left(\frac{10}{T_2 T_1}\right)} \quad [\text{Eqn. 3}]$$

Following rearrangement,

$$\ln Q_{10} = \left(\frac{E_a}{R}\right)\left(\frac{1}{T_1 T_2}\right) \quad [\text{Eqn. 4}]$$

Another essential biological constant derived by substituting the given values into the reaction rates equation regulated by the Q_{10} rule is the coefficient of temperature or theta (θ) value (simplified Arrhenius temperature coefficient);

$$kT = k_{20}\theta(T-20) \quad [\text{Eqn. 5}]$$

RESULT AND DISCUSSION

Bacterial growth rate is affected by temperature. Plotting \ln growth rate (per day) vs $1/T$ yielded a Chevron-like graph, revealing a discontinuous curve across the entire temperature range (Fig. 2). It was surprising to find a break point at 28.05 °C. At temperatures ranging from 20 to 27 °C, growth on SDS showed an activation energy of 41.72 kJ/mol, with a rise in activation energy to 84.72 kJ/mol at temperatures between 30 and 42 °C, as shown in regression analysis results in Table 1. As the temperature rises, the maximum rate of bacterial growth on SDS is found to be 27 °C, and the maximum rate drops as the temperature increases (Fig. 1).

On acrylamide, a previous study found an activation energy of 14.96 KJ/mol for *Pseudomonas* sp. strain DrYJ [16], which is definitely much lower. It was found that the activation energy obtained using the Arrhenius model was within the range of activation energies reported in the literature for diverse biodegradation processes (Table 2). Dismantling the connections appears to necessitate more effort. Energy savings can be achieved by raising the temperature. Many publications on activation energy computed from metabolic process rates at various temperatures report only one activation energy across a wide temperature range, rather than reporting on the occurrence of two activation energies. One study shows a higher activation energy at higher temperatures than at lower temperatures, whereas the other shows the opposite occurrence in the opposite direction (Table 2). It was found that the activation energy obtained using the Arrhenius model was within the range of activation energies reported in the literature for diverse biodegradation processes (Table 2). Dismantling the bonds appears to necessitate more effort. Energy savings can be achieved by raising the temperature. Many publications on activation energy computed from metabolic process rates at various temperatures report only one activation energy across a

wide temperature range, rather than reporting on the occurrence of two activation energies. One study shows a higher activation energy at higher temperatures than at lower temperatures, whereas the other shows the opposite occurrence in the opposite direction (Table 2). An example is the growth of *Bacillus* sp. JF8 on the xenobiotic polychlorinated biphenyl (PCB) with an activation energy from 20 to 46 °C of 12.1 Kj/mol and from 50 to 70 °C the activation energy is 31.4 Kj/mol [29]. A contrasting study shows the growth on phenol from 15 to 30 °C by *Pseudomonas* sp. AQ5-04 of 38.92 Kj/mol and from 35-45 °C, the activation energy is 11.34 Kj/mol [30].

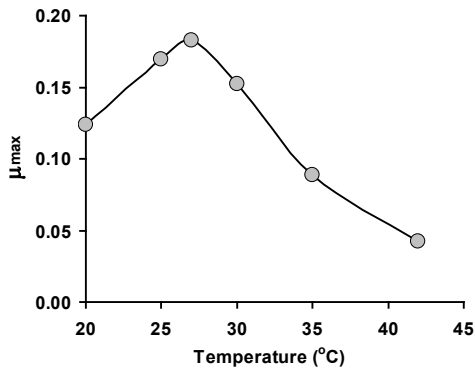


Fig 1. The effect of temperature on the specific growth rate of *Serratia marcescens* strain DRY6 on SDS.

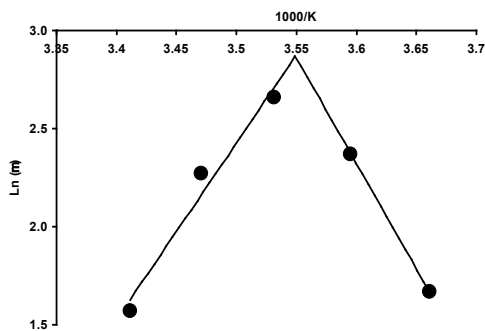


Fig 2. Growth rate of SDS by *Serratia marcescens* strain DRY6 in the form Arrhenius plot.

Table 1. Regression analysis was performed on the Arrhenius plot of the SDS biodegradation rate by *Serratia marcescens* strain DRY6.

Distribution of the experimental points	Three points to the left, three points to the right
Temperature range °C	Left part 30, 35 and 42
Regression equation	$y = 10.195x - 35.508$
Coefficient of determination	0.97
$\tan \alpha \pm$ Standard error	10.19 ± 0.05
$E_a \pm$ Standard error, kJ/mol	84.72 ± 0.41
t-Statistic	205.01
Degrees of freedom	2
Temperature range °C	Right part 20, 25 and 27
Regression equation	$y = -5.0207x + 15.043$
Coefficient of determination	2
$\tan \alpha \pm$ Standard error	-5.02 ± 0.47
$E_a \pm$ Standard error, kJ/mol	41.72 ± 3.89
t-Statistic	-10.72
Degrees of freedom	2
Intersection coordinates, (x, y)	Break points data 3.32, -1.642
Break point temperature °C	28.05

Table 2. Arrhenius temperature characteristics for growth on numerous xenobiotics.

Microorganisms	Temperature range (°C)	Substrate	DH* activation (Kj.mol ⁻¹)	apparent energy Ref
activated sludge	10–20	phenol	39.0	[31]
<i>Selanastrum capricornutum</i>	20–28	phenol	28.4	[32]
aerobic fluidized-bed reactors (FBRs)	14–16.5	2,4,6-trichlorophenol (TCP), 2,3,4,6-PCP tetrachlorophenol (TeCP), and pentachlorophenol (PCP)	TCP and TeCP 194 194 59-130	126- [33]
<i>Pseudomonas putida</i> Q5	10–25	phenol	61.6	[24]
Acclimated cultures	15–30	nonylphenol	42.7	[34]
<i>Pseudomonas putida</i> MTCC 1194	15–30	phenol	57.74	[35]
<i>Bacillus</i> sp. JF8	20–70	polychlorinated biphenyl (PCB)	12.1 (20–46 °C) 31.4 (50–70 °C)	[29]
<i>Pseudomonas</i> AQ5-04	sp. 15–45	phenol	38.92 (15–30 °C) 11.34 (35–45 °C)	[30]
<i>Pseudomonas</i> Strain DrYJ7	sp. 10–20	SDS	14.96	[16]
<i>Cupriavidus</i> strain CNP-8	sp. 20–40	2-chloro-4-nitrophenol	75.16 88.71	[36]
<i>Escherichia coli</i> BL21	20–50	Chromate	28.01	[37]
<i>Ochrobactrum intermedium</i>	25–35	Chromate	120.69	[38]
<i>Shewanella oneidensis</i> MR-1	25–40	Selenate	Control 62.90 TPPS-supplemented system 47.33	system [39]
anaerobic sludge	30–55	Reactive Red 2	22.9	[40]
activated bacterial consortium	20–37	Remazol Black B	48.8	[41]
<i>Enterobacter</i> strain (GY-1)	sp. 20–35	Reactive Black (RB 5)	5 35.56	[42]
<i>Escherichia coli</i> NO3	20–45	Reactive red 22	27.49	[43]
<i>Pseudomonas aeruginosa</i>	15–45	Reactive Black and Acid Red 360	39 RB39 61.89 360 AR360 81.18	[44]
<i>Pseudomonas</i> LPM-410	sp. 20–28	EDTA	91.2	[45],
<i>Pseudomonas</i> AQ5-04	sp. 15–45	phenol	38.92 (15–30 °C) 11.34 (35–45 °C)	[30]
<i>Cupriavidus</i> sp.	20–25 30–40	2-chloro-4-nitrophenol	88.71 (20-25) 75.16 (30-40)	[17]
<i>Pseudomonas</i> strain DrYJ7	sp. 10–20	Acrylamide	14.96	[16]

Note: (TPPS) Meso-tetrakis (4-sulfonatophenyl) porphyrin mediator

The temperature is an essential factor to take into account while thinking about bacteria. Nearly every aspect of the folding, structure, and stability of biomolecules, in addition to the rate at which metabolic reactions occur, are all influenced. The ability of microbes, such as pathogens, to sense the presence of a host and adapt their metabolic processes accordingly is essential to the microbe's ability to live on in its environment. In order to modify their gene expression in response to changes in temperature, organisms have created specialized sensing systems that are capable of detecting temperature fluctuations in an indirect manner. Both a heat shock and a cold shock can result in the aggregation of aggregated proteins, but only a cold shock can result in the accumulation of stopped ribosomes [14,15,46–51]. Molecular thermosensors could be built using molecular switches as the building blocks. Changes in molecular architecture that result in dysfunctional activity are one example of direct temperature sensing. Other examples include temperature-responsive regulatory proteins and alterations in lipid membrane integrity in relation to fluidity. It is impossible to exaggerate the

importance of temperature in the regulation of biological processes. Applications in the field of temperature-controlled biotechnology could also stand to profit from this method. Changes in temperature bring to a plethora of metabolic modifications, many of which are closely connected with one another. In light of this, models that illustrate how the progression of a biological process varies as a function of temperature are an absolute must if we want to simplify things and get a better understanding of how everything is connected. As early as the 1900s, Arrhenius provided a mathematical account of how biochemical reactions respond to high or low temperatures in the form of his "primal" temperature model [47–51].

Bacteria need more energy to break down more complex xenobiotics when the activation energy is higher. Several xenobiotic-degrading microbial species have activation energies within the ranges observed in this study for both temperature ranges, as shown in **Table 2**. In contrast, the activation energies of typical mesophilic bacteria range from 33.5 to 50.3 kJ/mol [52]. It's possible that one of the temperature ranges examined had a higher activation energy. The activation energy was found to vary with temperature in this study, rather than being a constant [53]. The model is useful as an observational model because it can't account for all of the simultaneous interactions among the various biological processes. Instead of thinking in terms of the activation energy used in chemical reactions, we should consider activation energy to be the total temperature response of a microorganism [54].

These drawbacks notwithstanding, the model continues to see widespread use. Temperature affects the activation energy of microbes, which has been demonstrated in a variety of settings, such as the decolorization of various colors by various organisms (**Table 2**). Although the exact mechanism causing the transformation is still a mystery, two hypotheses can account for it. When transitioning from one state to another, the water features change and a "bottleneck" theory suggests the same processes occur simultaneously and quickly [55]. Based on numerous observations of Arrhenius break point temperatures, the first hypothesis does not appear to be valid. There are many reasons why it is difficult to prove the "bottle-neck" hypothesis, according to the "bottle-neck" theory. Because of the effect of temperature on the cell membrane, it will also change [56]. Academics continue to believe in the "bottleneck" theory [25,57].

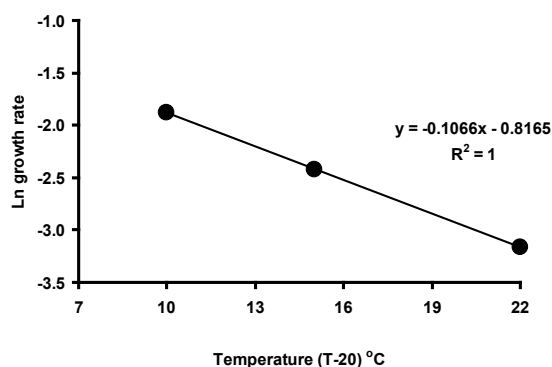


Fig. 3. Ln growth rate vs temperature plot for estimating theta.

It is also possible to calculate the Q_{10} values using the Arrhenius plots, or by monitoring growth rates at various incubation temperatures with ten degrees of variation [58]. The Arrhenius curve is the slope of the resultant plot when the bioreduction and growth rates are logarithmically plotted against

1000/temperature (Kelvin) (**Fig. 1**). For the temperature range of 30 to 42 °C, a Q_{10} value of 2.905 was obtained (**Fig. 3**). However, due to the dynamic nature of biological processes, many Q_{10} values may be found for each investigated temperature range. The conversion of molybdate to molybdenum blue yielded a 2.038 value [59]. A Q_{10} value of 2.31 was found in *Morganella* sp, yet another molybdenum reducer in nature. In order to properly attribute the growth process to a specific biological activity, this value is essential. For oil degradation in a beach gravel column, previous research found Q_{10} to be 2.7 [60]. Nevertheless, a Q_{10} value of 2.2 for microbiological process was reported in another study on decane and toluene-contaminated soil [61]. Degradation rates for other petrochemical compounds affected by salinity was reported to exhibit a Q_{10} value of 2.2 [62], while acrylamide degradation between 25 and 45 °C in immobilized bacterial systems reported a Q_{10} value of 2.8 [63].

Declining temperatures frequently lead to an increase in the Q_{10} value [64,65]. The Q_{10} value for the phenol degradation by *Pseudomonas* sp. strain AQ5-04 was 1.834 [30] while a Q_{10} value of 2.17 was calculated for the growth rate of this organism on SDS. A lower Q_{10} value of 2.17 is reported in another study on acrylamide biodegradation by the Antarctic bacterium *Pseudomonas* sp. strain DRYJ7 [16]. The theta value was determined to be 1.11 (**Fig. 3**), close to the theta value found of 1.08 for molybdenum reduction by *Serratia* sp. strain HMY1 [59]. A lower theta value of 1.03 is reported for the growth rate on SDS by the Antarctic bacterium *Pseudomonas* sp. strain DRYJ7 [16]. There are reports of xenobiotics being broken down with theta values of up to 16.2, but theta values are typically between 1.1 and 1.7, which is within the typical range for many biological processes [66]. In the biodegradation of nonylphenol, a theta value of 1.06 was observed [34].

CONCLUSION

SDS biodegradation requires an activation energy that can be seen in an Arrhenius plot with two activation energies. This is the first study to demonstrate this. Temperature has a significant impact on microbial growth and metabolic activity on their substrates. Because they are so small, microorganisms are extremely sensitive to temperature changes in their environment. An apparent activation energy graph was presented, with a break point at 28.05 °C showing the apparent activation of *Serratia marcescens* DRY6 on SDS. There were two different activation temperatures: 20–27 °C and 30–42 °C, both with activation energies of 41.72 and 84.72 kJ/mol, respectively, based on the regression analysis results. A Q_{10} value of 2.905 and a theta value of 1.11 were calculated for the studied temperature range (30–42 °C). More than three times as powerful as mesophilic bacteria at temperatures between 30 and 42 °C, the quantum. Breaking an amide bond is thought to require a much higher activation energy. Additional research and inquiry into the parameters themselves are being conducted in order to better understand the relationship between temperature and growth kinetics.

REFERENCES

- Pettersson A, Adamsson M, Dave G. Toxicity and detoxification of Swedish detergents and softener products. *Chemosphere*. 2000;41(10):1611–20.
- Rosety M, Ribelles A, Rosety-Rodriguez M, Carrasco C, Ordonez FJ, Rosety JM, et al. Morpho-histochemical study of the biological effects of sodium dodecyl sulphate on the digestive gland of the Portuguese oyster. *Histol Histopathol*. 2000;15(4):1137–43.
- Dhouib A, Hamad N, Hassaïri I, Sayadi S. Degradation of anionic surfactants by *Citrobacter braakii*. *Process Biochem*. 2003;38(8):1245–50.

4. Sharvelle SE, Garland J, Banks MK. Biodegradation of polyalcohol ethoxylate by a wastewater microbial consortium. *Biodegradation*. 2008;19(2):215–21.
5. Ambily PS, Jisha MS. Biodegradation of anionic surfactant, sodium dodecyl sulphate by *Pseudomonas aeruginosa* MTCC 10311. *J Environ Biol*. 2012;33(4):717–20.
6. Asok AK, Jisha MS. Biodegradation of the anionic surfactant linear alkylbenzene sulfonate (LAS) by autochthonous pseudomonas sp. *Water Air Soil Pollut*. 2012;223(8):5039–48.
7. Chaturvedi V, Kumar A. Diversity of culturable sodium dodecyl sulfate (SDS) degrading bacteria isolated from detergent contaminated ponds situated in Varanasi city, India. *Int Biodeterior Biodegrad*. 2011;65(7):961–71.
8. Chaturvedi V, Kumar A. Isolation of a strain of *Pseudomonas putida* capable of metabolizing anionic detergent sodium dodecyl sulfate (SDS). *Iran J Microbiol*. 2011;3(1):47–53.
9. Chaturvedi V, Kumar A. Presence of SDS-degrading enzyme, alkyl sulfatase (SdsA1) is specific to different strains of *Pseudomonas aeruginosa*. *Process Biochem*. 2013;48(4):688–93.
10. Rahman MF, Rusnam M, Gusmanizar N, Masdor NA, Lee CH, Shukor MS, et al. Molybdate-reducing and SDS-degrading *Enterobacter* sp. Strain Neni-13. *Nova Biotechnol Chim*. 2016 Dec 1;15(2):166–81.
11. Içgen B, Salik SB, Goksu L, Ulusoy H, Yilmaz F. Higher alkyl sulfatase activity required by microbial inhabitants to remove anionic surfactants in the contaminated surface waters. *Water Sci Technol J Int Assoc Water Pollut Res*. 2017 Nov;76(9–10):2357–66.
12. Payne WJ, Feisal VE. Bacterial utilization of dodecyl sulfate and dodecyl benzene sulfonate. *Appl Microbiol*. 1963;11:339–44.
13. Ohtake H, Fujii E, Toda K. Bacterial Reduction of Hexavalent Chromium: Kinetic Aspects of Chromate Reduction by *Enterobacter cloacae* HO1. *Biocatalysis*. 1990 Jan 1;4(2–3):227–35.
14. Ratkowsky DA, Ross T, McMeekin TA, Olley J. Comparison of Arrhenius-type and Bełehradek-type models for prediction of bacterial growth in foods. *J Appl Bacteriol*. 1991;71(5):452–9.
15. Minkevich IG, Satroudinov AD, Dedyukhina EG, Chistyakova TI, Kaparullina EN, Koshelev AV, et al. The effect of temperature on bacterial degradation of EDTA in pH-auxostat. *World J Microbiol Biotechnol*. 2006 Nov 1;22(11):1205–13.
16. Gafar AA, Manogaran M, Yasid NA, Halmi MIE, Shukor MY, Othman AR. Arrhenius plot analysis, temperature coefficient and Q_{10} value estimation for the effect of temperature on the growth rate on acrylamide by the Antarctic bacterium *Pseudomonas* sp. strain DRY17. *J Environ Microbiol Toxicol*. 2019 Jul 31;7(1):27–31.
17. Shukor MY. Arrhenius Plot Analysis of the Temperature Effect on the Biodegradation Rate of 2-chloro-4-nitrophenol. *Biog J Ilm Biol*. 2020 Dec 30;8(2):219–24.
18. Abubakar A. Arrhenius Plot Analysis, Temperature Coefficient and Q_{10} Value Estimation for the Effect of Temperature on the Rate of Molybdenum Reduction by *Serratia marcescens* strain DRY6. *J Environ Microbiol Toxicol*. 2021 Jul 31;9(1):21–6.
19. Abubakar A. Arrhenius Plot Analysis, Temperature Coefficient and Q_{10} Value Estimation for the Effect of Temperature on the Rate of Molybdenum Reduction by *Acinetobacter calcoaceticus* strain Dr Y12. *Bull Environ Sci Sustain Manag E-ISSN 2716-5353*. 2021 Jul 31;5(1):20–6.
20. Uba G, Yakasai HM, Babandi A, Ya'u M, Mansur A. The Effect of Temperature on the Rate of Molybdenum Reduction by *Enterobacter* sp. strain Dr.Y13: Arrhenius Plot Analysis, Temperature Coefficient and Q_{10} Value Estimation. *Bull Environ Sci Sustain Manag E-ISSN 2716-5353*. 2021 Jul 31;5(1):1–6.
21. Yakasai HM, Safiyanu AJ, Ibrahim S, Babandi A. Arrhenius Plot Analysis, Temperature Coefficient and Q_{10} Value Estimation for the Effect of Temperature on Molybdenum Reduction Rate by *Pantoea* sp. strain HMY-P4. *J Environ Microbiol Toxicol*. 2021 Jul 31;9(1):16–20.
22. Singh RK, Kumar S, Kumar S, Kumar A. Biodegradation kinetic studies for the removal of p-cresol from wastewater using *Gliomastix indicus* MTCC 3869. *Biochem Eng J*. 2008;40(2):293–303.
23. Reardon KF, Mosteller DC, Bull Rogers JD. Biodegradation kinetics of benzene, toluene, and phenol as single and mixed substrates for *Pseudomonas putida* F 1. *Biotechnol Bioeng*. 2000;69(4):385–400.
24. Onysko KA, Budman HM, Robinson CW. Effect of temperature on the inhibition kinetics of phenol biodegradation by *Pseudomonas putida* Q5. *Biotechnol Bioeng*. 2000 Nov 5;70(3):291–9.
25. Angelova B, Avramova T, Stefanova L, Mutafov S. Temperature effect on bacterial azo bond reduction kinetics: an Arrhenius plot analysis. *Biodegradation*. 2008;19(3):387–93.
26. Zwietering MH, de Koos JT, Hasenack BE, de Witt JC, van't Riet K. Modeling of bacterial growth as a function of temperature. *Appl Environ Microbiol*. 1991 Apr;57(4):1094–101.
27. Othman AR, Yusof MT, Shukor MY. Biodegradation of Sodium Dodecyl Sulphate (SDS) by *Serratia marcescens* strain DRY6. *Bioremediation Sci Technol Res*. 2019 Dec 28;7(2):9–14.
28. Arrhenius S. Über die Reaktionsgeschwindigkeit bei der Inversion von Rohrzucker durch Säuren. *Z Für Phys Chem*. 1889 Jan 1;
29. Mukerjee-Dhar G, Shimura M, Miyazawa D, Kimbara K, Hatta T. bph genes of the thermophilic PCB degrader, *Bacillus* sp. JF8: characterization of the divergent ring-hydroxylating dioxygenase and hydrolase genes upstream of the Mn-dependent BphC. *Microbiology*. 2005;151(12):4139–51.
30. Aisami A, Yasid NA, Johari WLW, Shukor MY. Estimation of the Q_{10} value; the temperature coefficient for the growth of *Pseudomonas* sp. aq5-04 on phenol. *Bioremediation Sci Technol Res*. 2017 Jul 31;5(1):24–6.
31. Benedek P, Farkas P. Influence of temperature on the reactions of the activated sludge process. In: Murphy RS, Nyquist D, Neff PW, editors. *Proceedings of the international symposium on water pollution control in cold climates*. University of Alaska, Washington, DC: Environmental Protection Agency; 1970.
32. Reynolds JH, Middlebrooks EJ, Procella DB. Temperature-toxicity model for oil refinery waste. *J Environ Eng Div*. 1974;100(3):557–76.
33. Melin ES, Jarvinen KT, Puhakka JA. Effects of temperature on chlorophenol biodegradation kinetics in fluidized-bed reactors with different biomass carriers. *Water Res*. 1998 Jan 1;32(1):81–90.
34. Jahan K, Ordóñez R, Ramachandran R, Balzer S, Stern M. Modeling biodegradation of nonylphenol. *Water Air Soil Pollut Focus*. 2008 Aug 1;8(3–4):395–404.
35. Bandyopadhyay SK, Chatterjee K, Tiwari RK, Mitra A, Banerjee A, Ghosh KK, et al. Biochemical studies on molybdenum toxicity in rats: effects of high protein feeding. *Int J Vitam Nutr Res*. 1981;51(4):401–9.
36. Bedade DK, Singhal RS. Biodegradation of acrylamide by a novel isolate, *Cupriavidus oxalaticus* ICTDB921: Identification and characterization of the acrylamidase produced. *Bioresour Technol*. 2018 Aug 1;261:122–32.
37. Guo J, Lian J, Xu Z, Xi Z, Yang J, Jefferson W, et al. Reduction of Cr(VI) by *Escherichia coli* BL21 in the presence of redox mediators. *Bioresour Technol*. 2012 Nov 1;123:713–6.
38. Kavita B, Keharia H. Reduction of hexavalent chromium by *Ochrobactrum intermedium* BCR400 isolated from a chromium-contaminated soil. *3 Biotech*. 2012 Mar;2(1):79–87.
39. Zhao R, Guo J, Song Y, Chen Z, Lu C, Han Y, et al. Mediated electron transfer efficiencies of Se(IV) bioreduction facilitated by meso-tetrakis (4-sulfonatophenyl) porphyrin. *Int Biodeterior Biodegrad*. 2020 Feb 1;147:104838.
40. dos Santos AB, Cervantes FJ, van Lier JB. Azo dye reduction by thermophilic anaerobic granular sludge, and the impact of the redox mediator anthraquinone-2,6-disulfonate (AQDS) on the reductive biochemical transformation. *Appl Microbiol Biotechnol*. 2004 Mar;64(1):62–9.
41. Dafale N, Wate S, Meshram S, Nandy T. Kinetic study approach of remazol black-B use for the development of two-stage anoxic-oxic reactor for decolorization/biodegradation of azo dyes by activated bacterial consortium. *J Hazard Mater*. 2008 Nov 30;159(2):319–28.
42. Chen G, Huang M hong, Chen L, Chen D hui. A batch decolorization and kinetic study of Reactive Black 5 by a bacterial strain *Enterobacter* sp. GY-1. *Int Biodeterior Biodegrad*. 2011 Sep 1;65(6):790–6.
43. Chang JS, Kuo TS. Kinetics of bacterial decolorization of azo dye with *Escherichia coli* NO3. *Bioresour Technol*. 2000 Nov 1;75(2):107–11.

44. Behzat B. Decolorization of Reactive Black 39 and Acid Red 360 by *Pseudomonas aeruginosa*. Water Sci Technol. 2015 Jul 6;72(8):1266–73.
45. Minkevich IG, Satroutdinov AD, Dedyukhina EG, Chistyakova TI, Kaparullina EN, Koshelev AV, et al. The effect of temperature on bacterial degradation of EDTA in pH-auxostat. World J Microbiol Biotechnol. 2006;22(11):1205–13.
46. Han MH. Non-linear Arrhenius plots in temperature-dependent kinetic studies of enzyme reactions: I. Single transition processes. J Theor Biol. 1972 Jun 1;35(3):543–68.
47. Ratkowsky DA, Olley J, McMeekin TA, Ball A. Relationship between temperature and growth rate of bacterial cultures. J Bacteriol. 1982;149(1):1–5.
48. Stannard CJ, Williams AP, Gibbs PA. Temperature/growth relationships for psychrotrophic food-spoilage bacteria. Food Microbiol. 1985;2(2):115–22.
49. McMeekin TA, Chandler RE, Doe PE, Garland CD, Olley J, Putro S, et al. Model for combined effect of temperature and salt concentration/water activity on the growth rate of *Staphylococcus xylosum*. J Appl Bacteriol. 1987;62(6):543–50.
50. Fernández A, Collado J, Cunha LM, Ocio MJ, Martínez A. Empirical model building based on Weibull distribution to describe the joint effect of pH and temperature on the thermal resistance of *Bacillus cereus* in vegetable substrate. Int J Food Microbiol. 2002;77(1–2):147–53.
51. Ucun H, Bayhan YK, Kaya Y. Kinetic and thermodynamic studies of the biosorption of Cr(VI) by *Pinus sylvestris* Linn. J Hazard Mater. 2008 May 1;153(1):52–9.
52. Tchobanoglous G, Schroeder ED. Water quality: Characteristics, modeling and modification. 1 edition. Reading, Mass: Pearson; 1985. 780 p.
53. Ratkowsky DA, Olley J, McMeekin TA, Ball A. Relationship between temperature and growth rate of bacterial cultures. J Bacteriol. 1982;149(1):1–5.
54. Melin ES, Ferguson JF, Puhakka JA. Pentachlorophenol biodegradation kinetics of an oligotrophic fluidized-bed enrichment culture. Appl Microbiol Biotechnol. 1997 Jun 1;47(6):675–82.
55. Kuhn HJ, Cometta S, Fiechter A. Effects of growth temperature on maximal specific growth rate, yield, maintenance, and death rate in glucose-limited continuous culture of the thermophilic *Bacillus caldolenax*. Eur J Appl Microbiol Biotechnol. 1980;10(4):303–15.
56. Ceuterick F, Peeters J, Heremans K, De Smedt H, Olbrechts H. Effect of high pressure, detergents and phospholipase on the break in the arrhenius plot of *Azotobacter nitrogenase*. Eur J Biochem. 1978;87(2):401–7.
57. Mutafov SB, Minkevich IG. Temperature effect on the growth of *Candida utilis* VLM-Y-2332 on ethanol. Comptes Rendus Acad Bulg Sci. 1986;39:71–4.
58. Funamizu N, Takakuwa T. Simulation analysis of operating conditions for a municipal wastewater treatment plant at low temperatures. In: Margesin R, Schinner F, editors. Biotechnological Applications of Cold-Adapted Organisms. Berlin, Heidelberg: Springer Berlin Heidelberg; 1999. p. 203–20.
59. Yakasai HM, Yasid NA, Shukor MY. Temperature Coefficient and Q_{10} Value Estimation for the Growth of Molybdenum-reducing *Serratia* sp. strain HMY1. Bioremediation Sci Technol Res. 2018 Dec 31;6(2):22–4.
60. Gibbs CF, Davis SJ. The rate of microbial degradation of oil in a beach gravel column. Microb Ecol. 1976 Mar 1;3(1):55–64.
61. Malina G, Grotenhuis JTC, Rulkens WH. The effect of temperature on the bioventing of soil contaminated with toluene and decane. J Soil Contam. 1999 Jul 1;8(4):455–80.
62. Oh YS, Kim SJ. Effect of temperature and salinity on the bacterial degradability of petroleum hydrocarbon. Korean J Microbiol Korea R. 1989;26(4):339–47.
63. Kim BY, Hyun HH. Production of acrylamide using immobilized cells of *Rhodococcus rhodochrous* M33. Biotechnol Bioprocess Eng. 2002 Aug 1;7(4):194.
64. Atlas RM, Bartha R. Fate and effects of polluting petroleum in the marine environment. In: Gunther FA, editor. Residue Reviews. Springer New York; 1973. p. 49–85. (Residue Reviews).
65. Deppe U, Richnow HH, Michaelis W, Antranikian G. Degradation of crude oil by an arctic microbial consortium. Extrem Life Extreme Cond. 2005 Dec;9(6):461–70.
66. Bagi A, Pampanin DM, Brakstad OG, Kommedal R. Estimation of hydrocarbon biodegradation rates in marine environments: A critical review of the Q_{10} approach. Mar Environ Res. 2013 Aug;89:83–90.