

## Isothermal Modeling and MOORA-Based Model Selection for Chromium Ion Adsorption by *Rhizopus nigricans*

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### Abstract

Adsorption of hexavalent chromium (Cr(VI)) onto *Rhizopus nigricans* biomass. The Langmuir model yielded a physically justified maximum adsorption capacity,  $Q_{mL}$ , of 112.756 mg/g (95 percent confidence interval from 98.780 to 126.731 mg/g), which almost matched the experimental maximum. Despite this, the uptake capacity increases continuously with equilibrium concentration, with the final data points showing a steady upward trend rather than leveling off into a distinct plateau. This indicates that complete saturation of the binding sites has not yet been reached within the tested concentration range. Because heterogeneous isotherm models better reflected the overall equilibrium profile, model selection based just on  $Q_m$  was insufficient. The isotherm equations were directly fitted using nonlinear regression, and the model's performance was assessed using several error criteria. According to MOORA approach, Fritz-Schlunder III, Redlich-Peterson, and the Toth model are the best models overall, better than Langmuir. Despite this, all these best models failed to produce maximum adsorption capacity values near the experimental value. It is suggested that Cr(VI) biosorption by *Rhizopus nigricans* biomass included heterogeneous binding sites rather than ideal uniform monolayer adsorption. Thus, the Langmuir isotherm model initially suggested to describe this monolayer adsorption process may be inaccurate, because the uptake capacity is still actively increasing, the true maximum saturation capacity is likely higher than this maximum recorded value. This behavior is most likely impacted by the chemical variety of the fungal surface. More equilibrium points and resampling methods for parameter validation should be included in future studies.

### INTRODUCTION

The cycling of chromium between its hexavalent Cr(VI) and trivalent Cr(III) forms and rapid industrial growth are the major causes of chromium pollution, which remains a major global environmental problem since chromium, particularly Cr(VI), is highly carcinogenic and toxic. Globally, chromium is primarily employed in electroplating, dye production, paper, steel fabrication, metallurgical processing, and leather tanning industries. These industries are the primary contributors to chromium pollution, resulting from the discharge of Cr(VI) due to inadequate waste management, fugitive emissions, and polluted effluents [1]. Hexavalent chromium toxicity in waterways and soils is represented by the highly mobile

chromate/dichromate anionic species. Hexavalent chromium is highly mobile and can easily bioaccumulate, endangering aquatic life and human health through inhalation, ingestion, and skin contact. Inconsistent monitoring and regulatory inadequacies are the main reason for continues chromium pollution, most especially in regions close to electroplating and mining industrial zones in developing countries [2]. In Malaysia, the major sources of chromium contamination include electroplating workshops, industrial estates and leather plants. These sources result from a combination of modern and historical pressures. Chromium contamination disproportionately affects peri-urban and rural communities by adding Cr(VI) and Cr(III) to surface waters and sediments through contaminated groundwater and downstream effluents [3]. River systems close to industrial clusters have been

found to have higher chromium concentrations in Malaysia. For maximum public health and ecosystem protection, there is a need for strong and enhanced effluent treatment, more integrated water quality management, and routine monitoring [3]. Chromium toxicity varies depending on the species. Trivalent chromium is less poisonous because of its limited solubility in water. Nevertheless, it is necessary in minimal amounts as an enzyme cofactor and becomes dangerous at high exposure levels. Conversely, Hexavalent chromium species can pass through cell membranes and have potent oxidative capacities. This species is hence extremely poisonous, mutagenic, and carcinogenic. Prolonged exposure to hexavalent chromium can cause respiratory issues, renal damage, and developmental impacts [1]. The endless need for chromium in corrosion-resistant coatings, steel alloys, and chemicals in tanning has resulted in a paradox: although chromium is required for the operation of many modern industries, excessive levels of it in the environment have serious negative effects on the environment. This issue has prompted researchers to move toward sustainable methods, such as more advanced treatment technologies and comprehensive, circular approaches, in order to control or clean up chromium contamination and waste [4].

Techniques for remediating chromium include precipitation, electrochemical treatment, chemical reduction, ion exchange, and bioremediation. Biosorption, the removal of Cr(VI) from contaminated environments by using biological materials (artificial or natural) as adsorbents, provides a number of important advantages, such as cost-effectiveness, the possibility of regenerating and reusing adsorbents, and compatibility with low-energy processes. Because of these benefits, biosorption is appealing in developing and poor nations [5–8]. The idea that biosorption is the most cost-effective and environmentally friendly method stems from a number of benefits: (1) it is economically feasible due to its low-cost feedstocks and simple processing requirements; (2) compared to physico-chemical methods, less energy is required; (3) it requires fewer chemical additives and less waste is produced; (4) there is possibility of recovering adsorbate and recycling of adsorbent, which allows for a circular use of chromium resources. Through biosorption, integrity of ecosystem is maintained, waste materials are utilized, and scalable solutions are provided in many places including Malaysia through decentralized treatment systems and pilot plants, these aligns with the UN's Sustainable Development Goals [9,10].

Dead *Rhizopus nigricans* biomass, such as *Rhizopus oryzae* and *Aspergillus niger*, have demonstrated encouraging chromium removal, especially in acidic environments where Cr(VI) species exhibit strong interactions with protonated biomass surfaces [11–14]. However, biosorption efficiency is affected by pH, particle size, biomass pretreatment, initial metal concentration, contact time, and competing ions. Thus, spent *Rhizopus nigricans* biomass should be considered a viable low-cost biosorbent; nonetheless, regeneration studies, column validation, and testing with actual industrial effluent are necessary for its practical application [15–17]. In this paper, different isotherms are used to represent the adsorption of chromium on *Rhizopus nigricans* biomass, and MOORA analysis is used to determine which model is the best.

## METHODS

### Data acquisition and fitting

The freeware Webplotdigitizer 2.5 was used to digitize Figure 3 data from a previously published work [18]. The program's digitizing capabilities have received praise for their dependability [19]. The data were then subjected to nonlinear regression using

Curve-Expert Professional (Version 2.6.5, copyright Daniel Hyams), a curve-fitting application. The implicit equations were solved using the MATLAB software suite (Mathworks, Massachusetts, USA).

## Isotherms

To avoid overfitting, the isothermal models used in this work (Table 1) are restricted to three-parameter models, which are displayed below.

**Table 1.** Mathematical models in the remodelling data.

Isotherm	$p$	Formula	Ref.
Henry's law	1	$q_e = HC_e$	[20]
Langmuir	2	$q_e = \frac{q_{mL}b_L C_e}{1 + b_L C_e}$	[21]
Jovanovic	2	$q_e = q_m(1 - e^{-K_j C_e})$	[22]
Freundlich	2	$q_e = K_f C_e^{\frac{1}{n_f}}$	[23]
Dubinin-Radushkevich <sup>a</sup>	2	Incorrect form $q_e = q_{mDR} \exp \left\{ -K_{DR} \left[ RT \ln \left( 1 + \frac{1}{C_e} \right) \right]^2 \right\}$ correct form $q_e = q_{mDR} \exp \left\{ -K_{DR} \left[ RT \ln \left( \frac{C_s}{C_e} \right) \right]^2 \right\}$	[24,25] [26,27]
Koble-Corrigan <sup>c</sup>	3	$q_e = \frac{AC_e^n}{1 + BC_e^n}$	[28]
Temkin <sup>a</sup>	3	$q_e = \frac{RT}{b_T} \{ \ln(a_T C_e) \}$	[29,30]
Redlich-Peterson	3	$q_e = \frac{K_{RP1} C_e}{1 + K_{RP2} C_e^{\beta_{RP}}}$	[31]
Sips <sup>c</sup>	3	$q_e = \frac{K_s q_{mS} C_e^{\frac{1}{n_s}}}{1 + K_s C_e^{\frac{1}{n_s}}}$	[32]
Toth	3	$q_e = \frac{q_m C_e}{(K_T + C_e^{n_T})^{1/n_T}}$	[33]
Hill <sup>c</sup>	3	$q_e = \frac{q_{mH} C_e^{n_H}}{K_H + C_e^{n_H}}$	[34]
BET	3	$q_e = \frac{q_{mBET} \alpha_{BET} C_e}{(1 - \beta_{BET} C_e)(1 - \beta_{BET} C_e + \alpha_{BET} C_e)}$	[35]
Vieth-Sladek	3	$q_e = \frac{q_{mVS} b_{VS} C_e}{(1 + b_{VS} C_e)^{n_{VS}}}$	[36]
Brouers-Sotolongo	3	$q_e = q_{mBS} \left[ 1 - \left( 1 + (0.5) \left( \frac{C_e}{T} \right)^\alpha \right)^{-2} \right]$	[37–39]
Fritz-Schlunder-III	3	$q_e = \frac{q_{mFS} K_{FS} C_e}{1 + K_{FS} C_e^{n_{FS}}}$	[40]
Fowler-Guggenheim <sup>a,b</sup>	3	$q_e = q_{mFG} \frac{K_L C_e e^{\frac{\alpha q_e}{q_{mFG}}}}{1 + K_L C_e e^{\frac{\alpha q_e}{q_{mFG}}}}$	[41]
Moreau	3	$q_e = q_{mM} \frac{b C_e + l b^2 C_e^2}{1 + 2b C_e + l b^2 C_e^2}$	[42]
Unilan	3	$q_e = \frac{q_{mU}}{2b_U} \ln \left( \frac{a_U + C_e e^{b_U q_e}}{a_U + C_e e^{-b_U q_e}} \right)$	

Note

<sup>a</sup>Models that include  $\ln(C_e)$ , implicit equation or any logarithmic function involving  $C_e$  will fail or produce NaN (undefined) results when  $C_e = 0$ . Therefore, data points where  $C_e = 0$  should be excluded from the analysis.

<sup>b</sup>Implicit equation or function.

<sup>c</sup>The Hill, Liu (omitted), Sips, and Koble–Corrigan isotherm models will be presented in their traditional forms to preserve their historical context and conceptual distinctions, despite their underlying equality [43,44].

## Statistical analysis

Statistical discriminatory or error function tests utilized in this study include corrected Akaike Information Criterion (AICc) [45,46], root-mean-squared error (RMSE), adjusted coefficient of determination ( $R^2$ ) [47], Bias Factor (BF), Accuracy Factor (AF) [48], Marquardt's percent standard deviation (MPSD) [49–51], and Bayesian Information Criterion [52]. *Obi* and *Pdi* are the

expected and observed values,  $n$  is the total number of observations, and  $p$  is the total number of model parameters [53].

RMSE was calculated using the following formula;

$$RMSE = \sqrt{\frac{\sum_{i=1}^n (Pd_i - Ob_i)^2}{n-p}} \quad (\text{Eqn. 1})$$

BF and AF were calculated using the following formula;

$$\text{Bias factor} = 10 \left( \sum_{i=1}^n \log \frac{(Pd_i / Ob_i)}{n} \right) \quad (\text{Eqn. 2})$$

$$\text{Accuracy factor} = 10 \left( \sum_{i=1}^n \log \frac{|(Pd_i / Ob_i)|}{n} \right) \quad (\text{Eqn. 3})$$

AICc was calculated using the following formula;

$$AICc = 2p + n \ln \left( \frac{RSS}{n} \right) + \frac{2(p+1)+2(p+2)}{n-p-2} \quad (\text{Eqn. 4})$$

BIC was calculated using the following formula;

$$BIC = n \ln \left( \frac{RSS}{n} \right) + k \ln(n) \quad (\text{Eqn. 5})$$

HQC was calculated using the following formula;

$$HQC = n \ln \left( \frac{RSS}{n} \right) + 2k \ln(\ln n) \quad (\text{Eqn. 6})$$

Adjusted coefficient of determination ( $R^2$ ) was calculated using the following formula;

$$\text{Adjusted } (R^2) = 1 - \frac{RMS}{S_e^2} \quad (\text{Eqn. 7})$$

$$\text{Adjusted } (R^2) = 1 - \frac{(1-R^2)(n-1)}{(n-p-1)} \quad (\text{Eqn. 8})$$

MPSD was calculated using the following formula;

$$MPSD = 100 \sqrt{\frac{1}{n-p} \sum_{i=1}^n \left( \frac{Ob_i - Pd_i}{Ob_i} \right)^2} \quad (\text{Eqn. 9})$$

### Application of Multiobjective Optimization by Ratio Analysis (MOORA) in Modeling

Since the best models frequently have a mixture of error function superiority, the Multiobjective Optimization by Ratio Analysis (MOORA) was used for multi-criteria decision-making (MCDM) in the modeling exercise. By concurrently assessing several performance criteria, this method makes it easier to choose the best model [54,55]. To achieve comparability across various performance indicators, the initial step in the process is to normalize the decision matrix. After that, the choice matrix was normalized. The following equation must be used for normalization because these measurements may have different units and magnitudes:

$$X'_{ij} = \frac{x_{ij}}{\sqrt{\sum_{i=1}^n x_{ij}^2}} \quad (\text{Eqn. 10})$$

Where  $X_{ij}$  is the original value of the  $j^{\text{th}}$  metric for the  $i^{\text{th}}$  model, and  $X'_{ij}$  is the normalized value.

### Ratio System Analysis

A ratio system approach was then used to aggregate the normalized values. The following formula was used to deduct non-beneficial criteria (the remaining error functions) or those that should be minimized from advantageous criteria (those that should be maximized,  $\text{adj}R^2$ ).

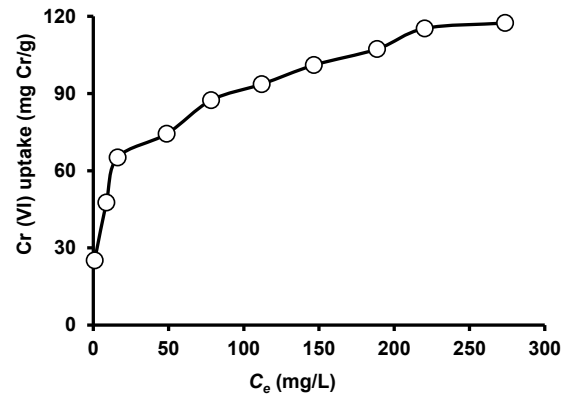
$$Y_i = \sum_{\text{beneficial}} X'_{ij} - \sum_{\text{non-beneficial}} X'_{ij} \quad (\text{Eqn. 11})$$

Where  $Y_i$  is the final score for the  $i^{\text{th}}$  model

It is advised to include weighted ratios in the analysis when some criteria were judged to be more important than others. Since there is currently no consensus in the literature over which of the error functions mentioned above should take precedence over the others, the recommendation to incorporate Weighted Ratios is not implemented. Ranking models according to their combined performance scores is the last stage. Better performance was reflected by higher scores. According to the specified decision criteria, the model with the highest value was deemed the best. This approach made it possible to compare kinetic models objectively and methodically, making it easier to determine which model performed the best while taking into account several performance criteria at once.

## RESULTS AND DISCUSSION

Nonlinear regression was used to apply various models to the equilibrium data of [18]. The experimental isotherm profile (**Fig. 1**) demonstrated that when the equilibrium concentration,  $C_e$ , increased, so did the *Rhizopus nigricans* biomass's uptake of Cr(VI). The adsorption capacity  $q_e$  increased dramatically from about 25 to 65 mg Cr/g at low  $C_e$ , suggesting a high initial affinity between Cr(VI) species and easily accessible binding sites on the *Rhizopus nigricans* biomass surface. At the highest tested concentration,  $q_e$  grew more gradually as  $C_e$  increased, reaching about 117 mg Cr/g. A fully horizontal plateau was not clearly created, despite the curve's inclination toward saturation, indicating that the true maximum adsorption capacity may not have been entirely reached within the examined range.



**Fig. 1.** Equilibrium adsorption of chromium to *Rhizopus nigricans* biomass raw data extracted using Webplot.

The adsorption behavior was better explained by heterogeneous and multi-parameter models than by more straightforward two-parameter equations, according to nonlinear isotherm fitting. Toth was ranked as the best model by MOORA, followed by Fritz-Schlunder III, Redlich-Peterson, Brouers-Sotolongo, and Sips, with Hill and Sips sharing the same position. Strong adsorption on a heterogeneous *Rhizopus nigricans* biomass surface with a variety of binding groups, including hydroxyl, carboxyl, phosphate, amine, and amide groups, is suggested by the good performance of Toth, Fritz-Schlunder III, and Redlich-Peterson [18,15–17]. Conversely, Temkin, Jovanovic, BET, Koble-Carrigan, and Henry performed worse, while Langmuir only came in at number twelve. Consequently, even though the curve indicated partial saturation, an ideal homogeneous monolayer model was unable to explain the adsorption process. Overall, Toth is the best model for Cr(VI) biosorption by the *Rhizopus nigricans* biomass, according to the MOORA data.

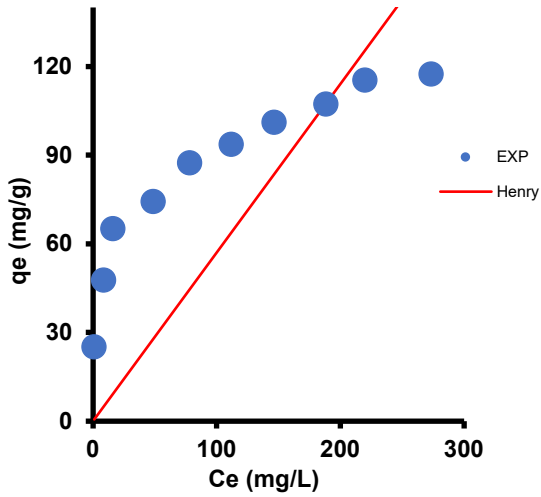


Fig. 2. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Henry model.

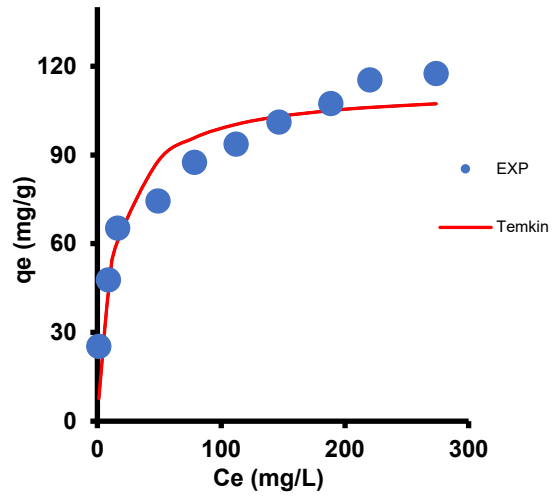


Fig. 5. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Temkin isotherm model.

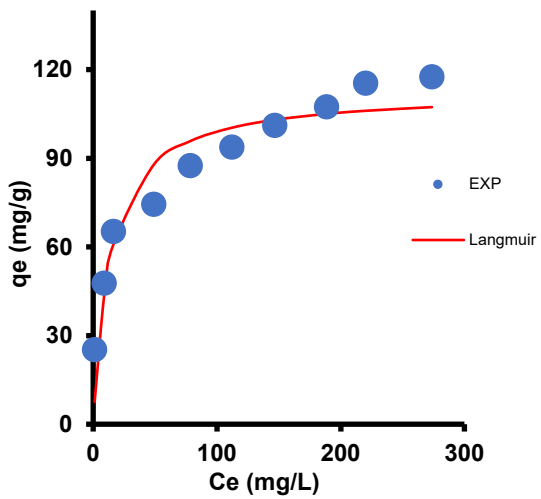


Fig. 3. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Langmuir isotherm model.

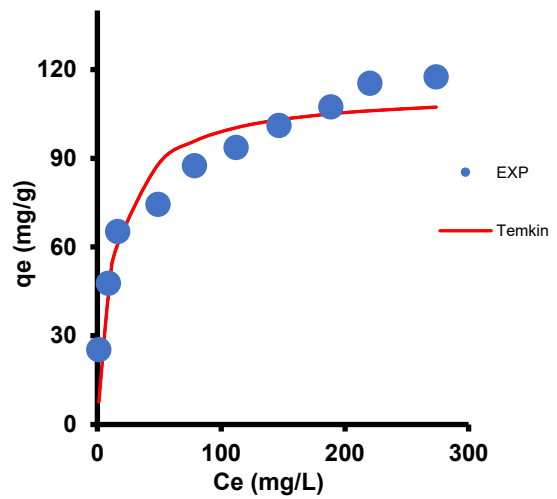


Fig. 6. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Koble-Corrigan isotherm model.

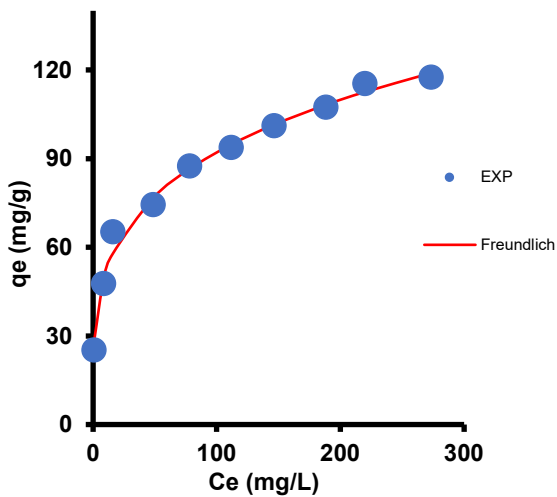


Fig. 4. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Freundlich isotherm model.

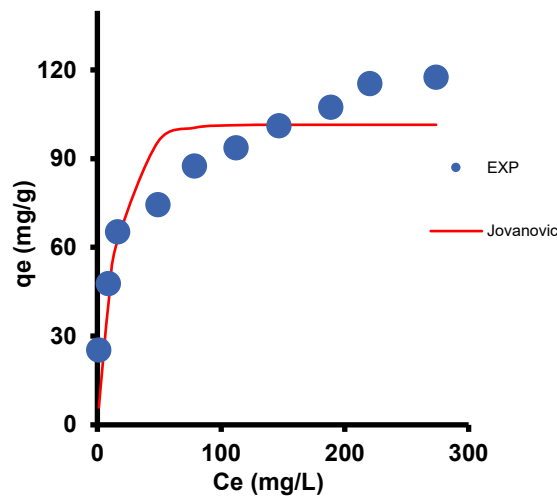


Fig. 7. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Jovanovic isotherm model.

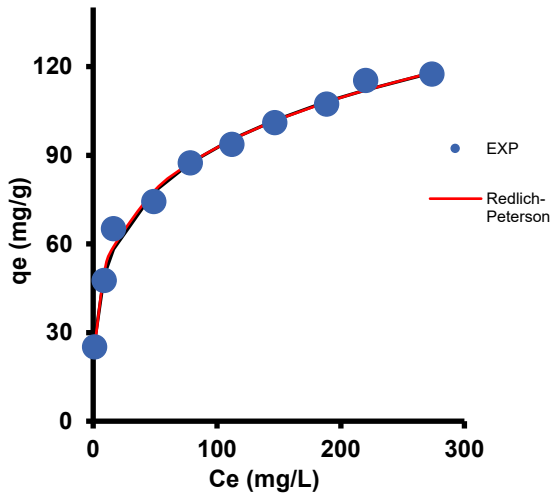


Fig. 8. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Redlich-Peterson isotherm model.

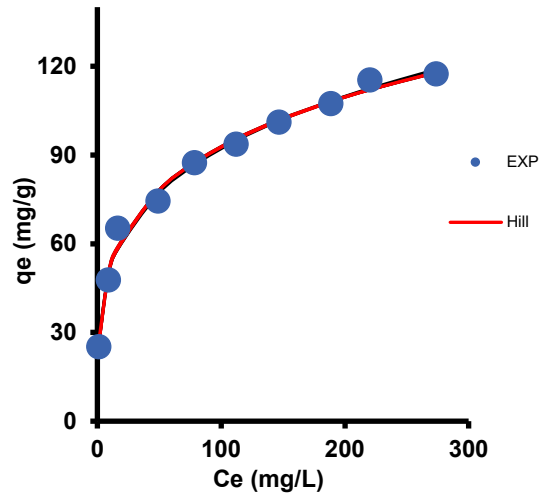


Fig. 11. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Hill isotherm model.

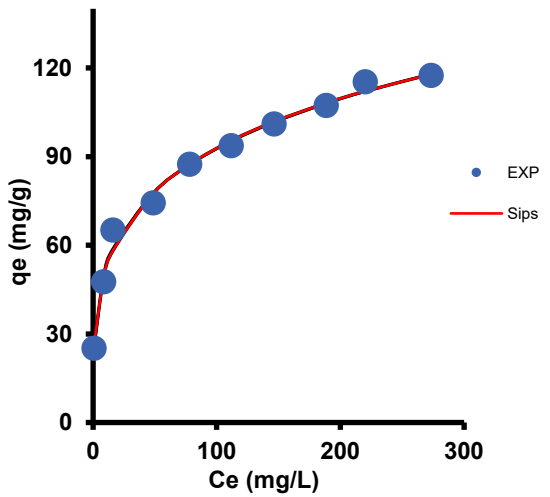


Fig. 9. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Sips isotherm model.

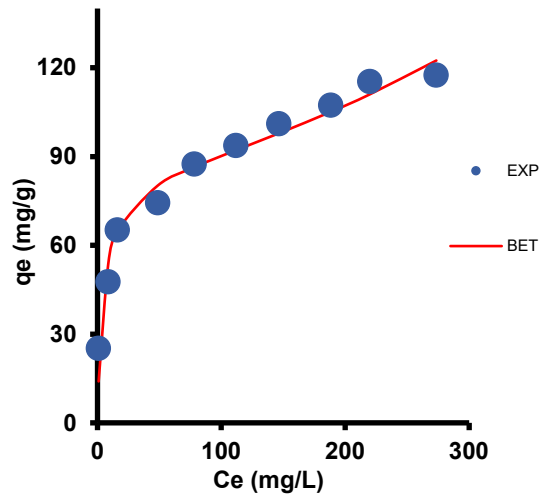


Fig. 12. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the BET isotherm model.

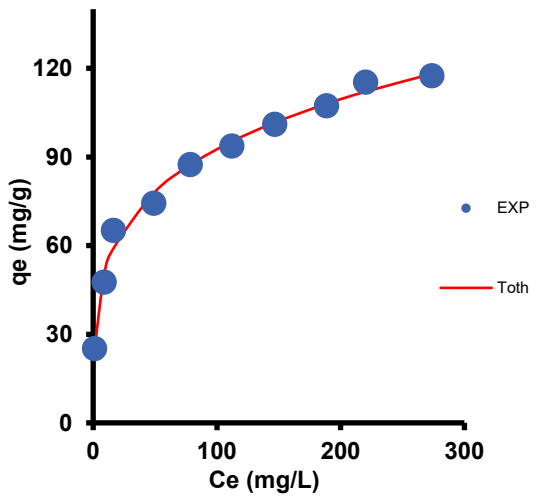


Fig. 10. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Toth isotherm model.

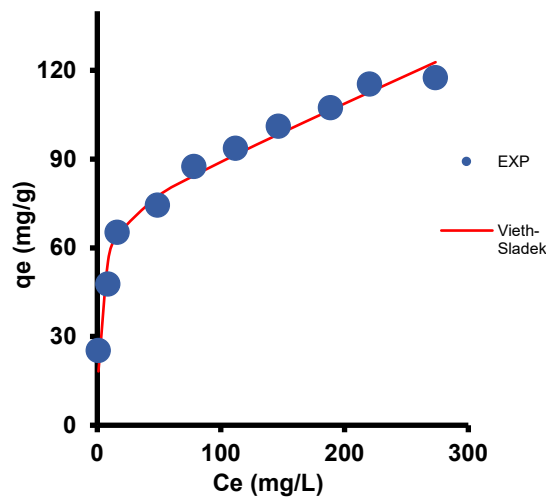


Fig. 13. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Vieth-Sladek isotherm model.

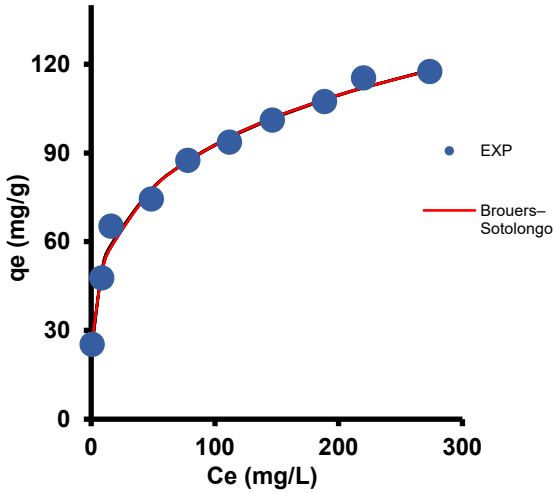


Fig. 14. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Brouers-Sotolongo isotherm model.

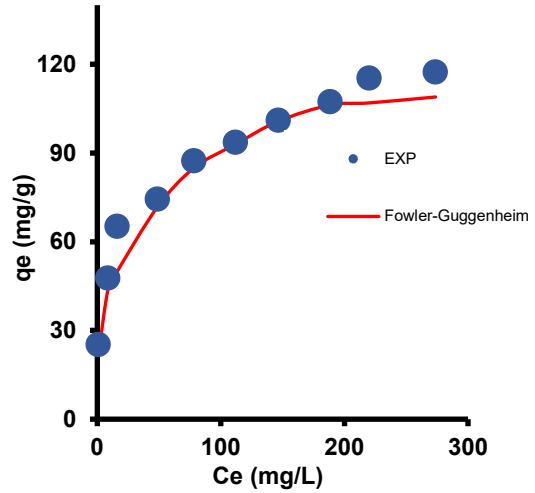


Fig. 17. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Moreau isotherm model.

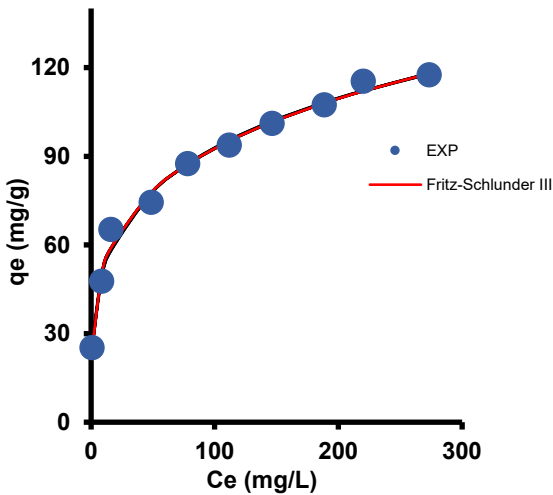


Fig. 15. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Fritz-Schlunder III isotherm model.

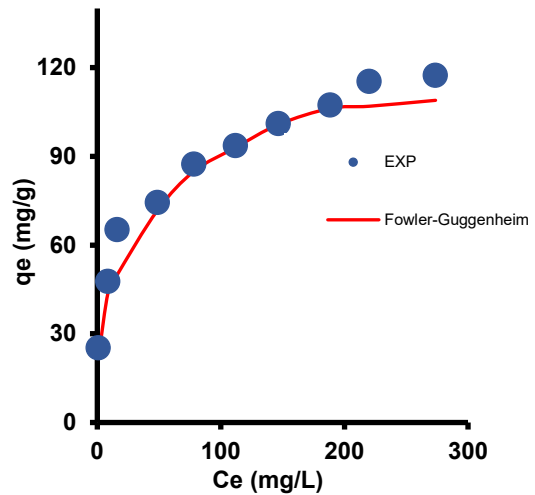


Fig. 18. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Fowler-Guggenheim isotherm model.

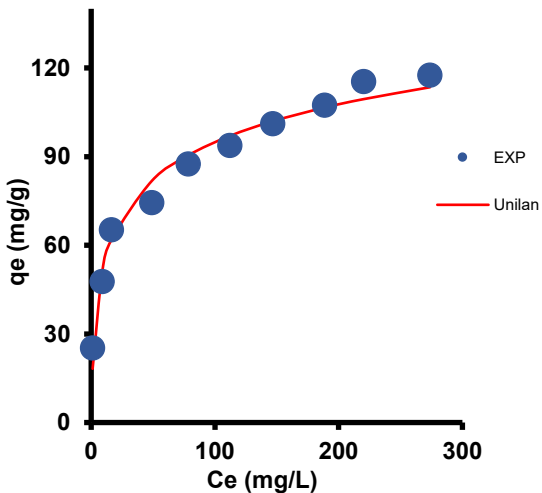


Fig. 16. Cr (VI) adsorption onto *Rhizopus nigricans* biomass modelled using the Unilan isotherm model.

Spent *Rhizopus nigricans* biomass is a desirable biosorbent for removing metal ions from aqueous effluents because it is frequently available as a waste by-product from fermentation or enzyme production operations, and it is also abundant and inexpensive. Spent or dead *Rhizopus nigricans* biomass can be employed in a wider range of environmental conditions, no nutrients requirement, and is less impacted by metal toxicity than living biomass. The biosorption capacity of *Rhizopus nigricans* biomass is primarily linked to its cell-wall constituents such chitin, glucans, proteins, and polysaccharides, which offer hydroxyl, carboxyl, phosphate, amine, and amide functional groups for metal binding. These groups can eliminate Cr(VI), Pb(II), Cd(II), Cu(II), Ni(II), and Zn(II) through complexation, ion exchange, surface precipitation, electrostatic attraction, and potential reduction of Cr(VI) to the less hazardous Cr(III).

**Table 2.** Error function analysis for the fitting of the isotherm of Cr (VI) adsorption onto *Rhizopus nigricans* biomass.

Model	<i>p</i>	MPSD	RMSE	<i>R</i> <sup>2</sup>	<i>adR</i> <sup>2</sup>	AICc	BIC	HQC	BF	AF
1 Henry	1	197.32	37.011	0.54	0.480	78.885	73.47	72.84	0.916	1.175
2 Langmuir	2	26.58	10.230	0.92	0.894	58.275	48.88	47.61	0.999	1.038
3 Freundlich	2	77.34	3.256	0.99	0.986	35.376	25.98	24.71	1.000	1.007
4 Temkin	3	29.12	10.936	0.92	0.877	66.274	51.18	49.28	0.999	1.038
5 Koble-Carrigan	2	315.29	3.140	0.96	0.95	34.651	25.26	23.99	1.001	1.007
6 Jovanovic	2	10.56	14.126	0.84	0.801	64.728	55.33	54.06	0.989	1.057
7 Redlich-Peterson	3	9.77	3.165	0.99	0.987	41.476	26.38	24.48	1.000	1.006
8 Sips	3	21.51	3.356	0.99	0.985	42.651	27.56	25.66	1.001	1.007
9 Toth	3	9.72	3.154	0.99	0.987	41.409	26.32	24.41	1.000	1.006
10 Hill	3	21.51	3.356	0.99	0.985	42.651	27.56	25.66	1.001	1.007
11 Khan	3	9.94	84.942	0.03	-0.449	#####	92.18	90.28	1.414	1.414
12 BET	3	173.39	6.067	0.97	0.958	54.489	39.40	37.49	0.992	1.016
13 Vieth-Sladek	3	9.94	5.044	0.98	0.969	50.797	35.71	33.80	0.993	1.016
14 Radke-Prausnitz	3	9.94	84.942	0.03	-0.449	#####	92.18	90.28	1.414	1.414
15 Brouers-Sotolongo	3	11.16	3.366	0.99	0.985	42.706	27.61	25.71	1.001	1.007
16 Fritz-Schlunder III	3	9.78	3.164	0.99	0.987	41.473	26.38	24.48	1.000	1.006
17 Unilan	3	29.61	5.302	0.98	0.965	51.796	36.70	34.80	0.999	1.017
18 Fowler-Guggenheim	3	54.58	4.395	0.98	0.976	48.043	32.95	31.05	0.998	0.998
19 Moreau	3	39.34	7.641	0.95	0.923	59.103	44.01	42.11	0.985	0.985

Note:  
 RMSE Root mean Square Error  
*adR*<sup>2</sup> Adjusted Coefficient of determination  
*p* no of parameters  
 AF Accuracy factor  
 BF Bias factor  
 BIC Bayesian Information Criterion  
 AICc Adjusted Akaike Information Criterion  
 HQC Hannan-Quinn information criterion

**Table 3.** Ranking of isothermal models based on MOORA.

No	Model	MOORA Score	Rank
1	Toth	0.05884	1
2	Fritz-Schlunder III	0.05775	2
3	Redlich-Peterson	0.05771	3
4	Brouers-Sotolongo	0.03546	4
5	Sips	0.01214	5
6	Hill	0.01214	5
7	Freundlich	-0.07768	7
8	Vieth-Sladek	-0.09553	8
9	Fowler-Guggenheim	-0.14920	9
10	Unilan	-0.15766	10
11	Moreau	-0.31211	11
12	Langmuir	-0.37354	12
13	Temkin	-0.43939	13
14	Jovanovic	-0.50297	14
15	BET	-0.54057	15
16	Koble-Carrigan	-0.63923	16
17	Henry	-1.55453	17

**Table 4.** Isothermal models' constants for Cr (VI) adsorption onto *Rhizopus nigricans* biomass.

Model	Parameter	Unit	Value	95% CI
Langmuir	<i>Q<sub>mS</sub></i>	(mg/g)	112.756	98.780 to 126.731
	<i>k<sub>L</sub></i>	(L/mg)	0.072	0.022 to 0.123
Redlich-Petersen	<i>k<sub>RP1</sub></i>	(L/g)	133.188	-149.374 to 415.749
	<i>k<sub>RP2</sub></i>	(L/mg)/g	4.248	-5.536 to 14.033
	<i>B<sub>RP</sub></i>	dimensionless	0.763	0.719 to 0.808
Toth	<i>q<sub>mT</sub></i>	(mg g <sup>-1</sup> )	30.94	24.473 to 37.410
	<i>K<sub>T</sub></i>	(mg/L) <sup>n<sub>T</sub></sup>	0.26	-0.330 to 0.850
	<i>n<sub>T</sub></i>	dimensionless	0.87	0.849 to 0.896
Fritz-Schlunder III	<i>q<sub>mFS</sub></i>	(mg g <sup>-1</sup> )	31.350	24.124 to 38.576
	<i>K<sub>FS</sub></i>	dimensionless	4.248	-5.534 to 14.031
	<i>n<sub>FS</sub></i>	dimensionless	0.763	0.719 to 0.808
Brouers-Sotolongo isotherm	<i>q<sub>mBS</sub></i>	mg g <sup>-1</sup>	311.586	-473.516 to 1096.688
	<i>K<sub>BS</sub></i>	mg <sup>-1/n<sub>BS</sub></sup> L <sup>1/n<sub>BS</sub></sup>	0.091	0.132 to 0.314
	<i>n<sub>BS</sub></i>	dimensionless	3.391	1.824 to 4.958
Hill isotherm	<i>q<sub>mH</sub></i>	mg g <sup>-1</sup>	493.692	-853.999 to 1841.382
	<i>n<sub>H</sub></i>	dimensionless	0.302	0.145 to 0.459
	<i>K<sub>H</sub></i>	dimensionless	17.330	-29.440 to 64.100
Sips	<i>q<sub>mS</sub></i>	mg g <sup>-1</sup>	493.135	-848.289 to 1834.559
	<i>K<sub>S</sub></i>	(L mg <sup>-1</sup> ) <sup>n<sub>S</sub></sup>	0.058	-0.097 to 0.213
	<i>n<sub>S</sub></i>	dimensionless	3.314	1.589 to 5.039

The Langmuir model calculated a maximum adsorption capacity, *Q<sub>mL</sub>*, of 112.756 mg/g with a standard error of 6.060 and a 95 percent confidence interval of 98.780 to 126.731 mg/g based on the adjusted Langmuir parameters. The Langmuir model gave a physically reasonable estimate of the upper adsorption capacity within the examined concentration range, as evidenced by the fact that this value is near the greatest experimentally observed adsorption capacity. With a 95 percent confidence interval of 0.022 to 0.123 L/mg, the Langmuir affinity constant, *K<sub>L</sub>*, was 0.072 L/mg, indicating a moderate affinity between Cr(VI) species and the *Rhizopus nigricans* biomass surface.

However, the Langmuir model should not be immediately regarded as the best model based only on its capacity estimate, even though the *Q<sub>mL</sub>* value is physically feasible. The Toth model provided the best overall fit, according to the MOORA ranking, followed by Fritz-Schlunder III and Redlich-Peterson. This should be expected because fungus has various functional groups, such as hydroxyl, carboxyl, phosphate, amine, and amide groups, in its cells that have chemical binding sites. This implies that an exact description of the entire equilibrium profile is provided by the models that account for surface heterogeneity and non-ideal adsorption.

The maximum computed chromium adsorption capacity of the Toth model *q<sub>mT</sub>* was 30.94 mg/g with an interval of 24.473 to 37.410 mg/g at 95 percent confidence. Despite having a lower value than the Langmuir *Q<sub>mL</sub>*, the Toth model was ranked best according to the total model-selection criteria. With a narrow 95 percent confidence interval of 0.849 to 0.896, its heterogeneity exponent, *n<sub>T</sub>*, was 0.87, showing a shift from perfect homogenous Langmuir-type adsorption. Consequently, the Toth model's ability to explain the equilibrium data is its major strength.

The Fritz-Schlunder III model provided a similar calculated maximum equilibrium adsorption capacity, *q<sub>mFS</sub>*, of 31.350 mg/g with a 95 percent confidence interval of 24.124 to 38.576 mg/g. With a confidence interval of 0.719 to 0.808 and exponent *n<sub>FS</sub>* of 0.763, non-ideal adsorption behavior is demonstrated again. The affinity-related parameter *K<sub>FS</sub>*, however, displayed a large confidence range that crossed zero, suggesting a lower level of

precision in this parameter estimation. Fritz-Schlunder III may therefore be helpful as an empirical fitting model, although care should be used when interpreting its parameters. With a 95 percent confidence interval of 0.719 to 0.808 and an estimated  $B_{RP}$  of 0.763, the Redlich-Peterson model likewise demonstrated heterogeneous adsorption. The model did not reduce to the Langmuir form and did not produce a finite theoretical  $Q_m$  because  $B_{RP}$  was obviously below unity. Rather, the experimental concentration range continues to influence the expected adsorption capacity. Furthermore, there was significant uncertainty in these fitted constants as  $k_{RP1}$  and  $k_{RP2}$  displayed large confidence ranges that crossed zero.

The three other models, namely Brouers–Sotolongo, Hill and Sips, showed higher estimated maximum adsorption values but with a wide range of confidence intervals, and in some cases, extended into negative lower limits. This wide range indicates that the maximum-capacity parameters were inappropriately covered by the available data, with the most likely reason being insufficient high-concentration data, as can be seen in **Fig. 1**, with no clear definition of a saturation plateau. The results imply that Cr(VI) adsorption onto the *Rhizopus nigricans* biomass was best explained by adsorption to a heterogeneous system. Due to this, the estimated  $Q_m$  values from Brouers–Sotolongo, Hill and Sips are not true theoretical maximum adsorption capacities but is likely model-dependent apparent capacities. More data are needed to describe the plateau region.

Overall, the adjusted Langmuir parameters demonstrate that Langmuir generated a reasonable and tightly controlled  $Q_{mL}$  of roughly 113 mg/g. Toth, Fritz-Schlunder III, and Redlich-Peterson's higher MOORA ranking, however, indicates that Cr(VI) adsorption onto the *Rhizopus nigricans* biomass was more accurately characterized as a heterogeneous process than as ideal monolayer adsorption on a uniform surface. Therefore, Langmuir may be useful for reporting an apparent maximum adsorption capacity, whereas Toth is more appropriate for describing the overall equilibrium behaviour.

In addition to standardizing performance ratings across criteria, MOORA is known for its computational efficiency and simplicity, which eliminates the need for complex pairwise comparisons or distance calculations. By managing criteria with different scales and units through organized normalization, MOORA allows for quick and easy evaluation. MCDM has been effectively and extensively applied in a number of model ranking tasks, such as the evaluation of Software Reliability Growth Models (SRGMs) [56]. Nevertheless, it is still rarely used in the selection of adsorption models [44]. As previously indicated, a major drawback of nonlinear modeling is its inability to adequately represent adsorption behavior of small datasets, like the eight data points utilized in this work. Instead of recognizing authentic adsorption patterns, an insufficient dataset increases the chances of recording random noise. This usually results in poor parameter estimation and overfitting of adsorption models, as well as a wrong comprehension of adsorption mechanisms. Small datasets reduce statistical power and widen confidence intervals. To solve issues related to small datasets and gain additional understanding of adsorption behavior, resampling techniques like bootstrapping, Monte Carlo simulations, and sensitivity analysis can be used to improve model resilience [57].

### Toth

The Toth isotherm, a three-parameter model that preserves the appropriate low- and high-concentration limits, describes adsorption on heterogeneously energetic surfaces. Langmuir isotherm is provided by  $n_T$  value approaching unity in the

equation, where  $n_T \rightarrow 1$  indicates stronger homogeneity. It retains physical meaning over wide ranges and reproduces Henry's rule for low  $C_e$  values. In several studies, the Toth isotherm usually outperforms Sips and Langmuir in the adsorption of organics and dyes, suggesting the model's ability to manage broad site-energy distributions and, at the same time, maintain the monolayer capacity approach [58,59].

### Redlich–Peterson model

The Redlich-Peterson (RP) isotherm is a flexible three-parameter equation that is helpful for heterogeneous surfaces because it interpolates between Langmuir and Freundlich behavior; in the model, the heterogeneity exponent  $\gamma$  ranges from 0 to 1; when  $\gamma = 1$ , the RP isotherm decreases to Langmuir with an apparent monolayer capacity. However, it adheres to Henry's law at low  $C_e$ . Nonlinear regression should be carried out over the linearized forms because of the association between its two parameters. The model is highly recommended and acclaimed as a benchmark for the three-parameter model [60–65]. The Redlich-Peterson model does not yield a finite theoretical  $Q_m$  because the fitted exponent, BRP, was 0.763 rather than 1.0. The predicted capacity is still variable on concentration since the equation does not converge to a stable saturation plateau when  $BRP < 1$ . Within the studied experimental range, the model predicted an apparent maximum uptake of approximately 118 mg/g at the highest equilibrium concentration, which is consistent with the observed upper  $q_e$  region. Consequently, rather than serving as a source of a conclusive monolayer  $Q_m$ , the Redlich-Peterson finding should be seen as an empirical description of heterogeneous adsorption [31,61,66].

### Langmuir isotherm

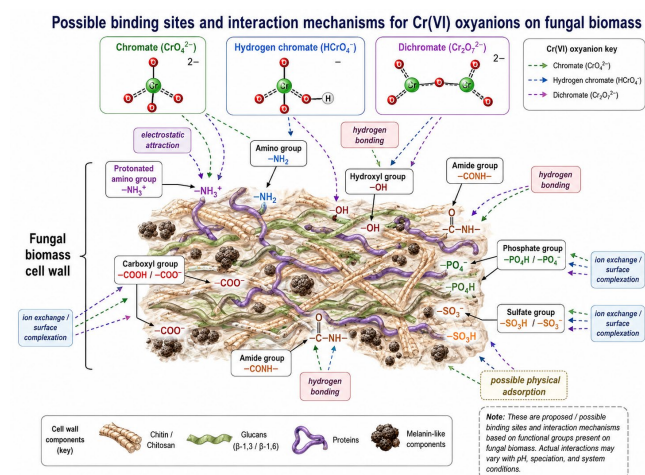
The Langmuir isotherm characterizes monolayer adsorption of an adsorbate onto a homogenous adsorbent surface, which only happens when the adsorbent has a uniform structure. In the model, all adsorption sites are identical and have the same energy [67]. Although the isotherm model and Henry's law are different, their underlying ideas are similar. The concept of monolayer coverage is supported by the theory that once a single layer of adsorbate molecules completely covers the surface, no more adsorption may take place at those locations [61]. The model has proven to be the most effective in simulating equilibrium chromium adsorption to adsorbents [68–81].

### Fritz-Schlunder III Model

Multilayer adsorption and interactions between adsorbed molecules on heterogeneous surfaces are explicitly described by the Fritz-Schlunder III model. With  $q_{mFS}$ ,  $K_{FS}$ , and  $n_{FS}$  as constants that represent the maximal adsorption capacity, the affinity of the binding sites, and the interactions between adsorbed molecules, respectively, this model takes into account the potential for multilayer adsorption at higher concentrations. Compared to the Langmuir model, the model seeks to offer a more adaptable and universal approach to adsorption isotherms that can explain a greater variety of adsorption processes. Multilayer adsorption and interactions between adsorbed molecules are explicitly taken into account by the Fritz-Schlunder III model, which incorporates a quadratic term in the denominator. The Fritz-Schlunder III model is particularly useful for systems where multilayer adsorption and molecule-molecule interactions are important.

The proposed binding sites and interaction mechanisms for Cr(VI) oxyanions on fungal biomass cell wall shows a multitude of functional groups interacting with the chromate anions (**Fig. 19**). In aqueous system, Cr(VI) occurs as chromate, hydrogen chromate, or dichromate species depending on solution pH and

the chromium concentration. The interactions of these oxyanions with fungal cell wall components such as chitin/chitosan, glucans, proteins, and melanin-like materials occur through several possible mechanisms. This mechanism include electrostatic attraction to protonated amino groups, hydrogen bonding with hydroxyl and amide groups, surface complexation with carboxyl or ion exchange, phosphate, and sulfate groups. In addition, physical adsorption within the heterogeneous cell wall matrix may occur. The following diagram highlights the possible Cr(VI) biosorption through the combined contribution of multiple functional groups. These dominant mechanism depends on pH, Cr(VI) speciation, biomass surface charge, and experimental conditions.



**Fig. 19.** Proposed Cr(VI) oxyanion binding to fungal biomass via electrostatic attraction, hydrogen bonding, ion exchange, surface complexation, and physical adsorption involving amino, hydroxyl, carboxyl, amide, phosphate, sulfate, and cell wall components (Illustration generated with the assistance of ChatGPT (OpenAI) based on the authors' instructions).

## CONCLUSION

Cr(VI) adsorption onto *Rhizopus nigricans* biomass was better characterized by a heterogeneous biosorption process with partial Langmuir-like saturation behavior, according to nonlinear modeling. With a 95 percent confidence interval of 98.780 to 126.731 mg/g, the corrected Langmuir model produced a physically meaningful  $Q_{ML}$  of 112.756 mg/g, which was near the maximum experimental adsorption capacity of roughly 117 mg/g. This implies that Langmuir can still be used to determine the apparent maximal uptake capacity. However, the MOORA assessment showed that heterogeneous isotherms offered a better explanation for the whole equilibrium profile, with the Toth model being the best overall model, followed by Fritz-Schlunder III and Redlich-Peterson. This is in line with the complex chemical structure of *Rhizopus nigricans* biomass, which contains many functional groups for metal binding. Even if MOORA and nonlinear regression improved model selection, the limited dataset remains a disadvantage. Future studies should incorporate additional equilibrium points, and more reliable parameter validation should be achieved by bootstrapping or Monte Carlo simulation.

## DECLARATION OF COMPETING INTEREST

The authors declare no competing financial interests or personal relationships.

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## AUTHORS' CONTRIBUTION

I.H. contributed to conceptualization, methodology, analysis, and writing the original draft. M.E.K. was involved in methodology, investigation, and data analysis. N.A.Y. contributed to investigation, resources, and data collection. A.A.B. participated in investigation, resources, and image generation. M.Y.S. contributed to methodology, validation, and writing review and editing.

## DATA AVAILABILITY STATEMENT

Data available on request

## AI USAGE DECLARATION

During the preparation of this work, the authors used generative AI technologies solely for language editing and refinement. All of the authors reviewed and edited the content. The authors take full responsibility for the final content of the publication.

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